Small Column Ion Exchange Testing of SuperLig® 644 for Removal of <sup>137</sup>Cs from Hanford Waste Tank 241-AP-101 Diluted Feed (Envelope A)

SK Fiskum ST Arm DL Blanchard, Jr BM Rapko

September 2002

Prepared for Bechtel National Inc. under Contract No. 24590-101-TSA-W0000-0004

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Battelle Pacific Northwest Division Richland, Washington 99352

# Completeness of Testing

This report describes the results of work and testing specified by TSP-W375-00-00028, Rev. 1 and CHG-TP-41500-013, Rev. 0 and TP-RPP-WTP-013, Rev. 0. The work and any associated testing followed the quality assurance requirements outlined in the Test Specification/Plan. The descriptions provided in this test report are an accurate account of both the conduct of the work and the data collected. Test plan results are reported. Also reported are any unusual or anomalous occurrences that are different from expected results. The test results and this report have been reviewed and verified.

Approved:	
Gordon H. Beeman, Manager WTP R&T Support Project	Date
G. Todd Wright, Manager Research and Technology	Date

# **Summary**

The River Protection Project-Waste Treatment Plant (RPP-WTP) baseline process for <sup>137</sup>Cs removal from Hanford high-level tank waste is ion exchange. The current flowsheet includes the use of Cs-selective, organic ion exchanger SuperLig<sup>®</sup> 644 (SL-644) material for Cs removal from the aqueous waste fraction. This material has been developed and supplied by IBC Advanced Technologies, Inc., American Fork, UT. The RPP-WTP Development Requirements Document (DRD)<sup>1</sup> task 8.2.6 and the RPP-WTP Research and Technology schedule<sup>2</sup> identify Cs and Tc ion exchange process verification tests [WTP Request for Proposal reference Standard 2 item (a)(3)(ii)].

Battelle Pacific Northwest Division (PNWD) was contracted to perform Cs ion exchange studies under contract 24590-101-TSA-W000-0004, and work breakdown structure BN.02.08.05. The Cs ion exchange activities are further defined in Technical Scoping Statement B-42a, which is included in Appendix C of the *Research and Technology Plan*<sup>2</sup>. These studies are to verify design and operating parameters for plant-scale ion exchange systems. Test results will also be used to validate ion exchange models.

#### **Objectives**

The Cs ion exchange test objectives were to develop load and elution breakthrough profiles using AP-101 diluted feed (AP-101DF); produce and characterize the Cs eluate for use in evaporation tests; remove  $^{137}$ Cs from the AP-101DF to meet LAW vitrification criteria; and develop batch distribution coefficient values for AP-101DF. The final effluent was to contain <1.75E-5 Ci  $^{137}$ Cs per mole Na, equivalent to < 0.087  $\mu$ Ci  $^{137}$ Cs/mL, based on a 20 wt% waste Na<sub>2</sub>O loading in the waste glass. Direction from BNI for calculating Na<sub>2</sub>O loading was later reduced to 14 wt%. All testing objectives were met.

#### **Conduct of Test**

This report summarizes testing of the SL-644 in batch-contact studies and in a dual small-column system. The test matrix was Hanford tank waste 241-AP-101 (Envelope A) diluted to 5 M Na $^+$  (AP-101DF). Batch contacts were performed with the waste at three Cs concentrations with a phase ratio of 100 (liquid volume to exchanger mass ratio) with both SL-644 and crystalline silicotitanate (IE-911, an alternative Cs ion exchanger). Ion exchange processing was conducted with two small columns in series with resin bed volumes (BVs) of 10.6 mL (L/D = 4.4) during the conditioning phase with 0.25 M NaOH, and 9.7 mL (L/D = 4) during the AP-101DF loading phase. Proper functioning of the ion exchange apparatus and resin beds had initially been tested with an AW-101 simulant. Approximately 1.2 L, equivalent to 112 BVs of AP-101DF waste were processed. All ion exchange process steps were tested

<sup>&</sup>lt;sup>1</sup> PL-W375-TE00002, Rev. 1, River Protection Project Waste Treatment Plant Development Requirements Document, October 31, 2000, M. E., Johnson and T. W. Crawford, CH2MHill Hanford Group, Inc., Richland, WA. DRAFT.

<sup>&</sup>lt;sup>2</sup> Barnes S, R Roosa, and R Peterson. 2002. *Research and Technology Plan.* 24590-WTP-PL-RT-01-002 Rev. 1, RPP-WTP project.

including resin bed preparation, loading, feed displacement, water rinse, elution, eluant rinse, and resin regeneration.

# **Results and Performance Against Objectives**

The batch-contact performance data are summarized in Table S1. The Cs  $\lambda$  value (column distribution ratio) represents a measure of the effective capacity of the SL-644 resin; the higher the  $\lambda$  value, the higher loading capacity. Results from the SL-644 batch-contact tests, resulted in an estimated batch distribution (K<sub>d</sub>) value of 690 mL/g in the feed condition (Na/Cs mole ratio of 1.4 E+5). Using a bed density of 0.230 g/mL in the 0.25 M NaOH condition, the Cs  $\lambda$  value is predicted to be 160 BVs. The IE-911 feed condition K<sub>d</sub> value was interpolated to be 710 mL/g. With a bed density of 1.04 g/mL, the estimated  $\lambda$  value is 740 BVs.

Flo	<b>ow</b>	Extrapola Cs break B	through,			K <sub>d</sub> , mL/g (feed condition)		Predicted Cs λ, BV	
	te //h)	Lead column	Lag column	Composite DF <sup>(1)</sup>	Maximum DF <sup>(2)</sup>	SL-644	IE-911	SL-644	IE-911
3	3	190	_	7.78E+5	1.4E+6	690	710	160	740

**Table S1.** Summary of Performance Measures

- (1) The decontamination factor was calculated by dividing the feed Cs concentration by the composite effluent Cs concentration.
- (2) The maximum DF was obtained by dividing the feed Cs concentration by the lowest lag column sample Cs concentration.

The ion exchange column processing performance data are also summarized in Table S1. The maximum ion exchange Cs breakthrough obtained was 0.27% at 112 BVs (regeneration condition). Based on the AP-101DF column load data, the lead column 50% Cs breakthrough was extrapolated to be 190 BVs (resin BV in the 0.25 M NaOH regeneration condition). A 50% Cs breakthrough value for the lag column could not be estimated; virtually no measurable Cs broke through the lag column. The decontamination factor (DF) for  $^{137}$ Cs, listed in Table S1, was based on  $^{137}$ Cs concentration in the feed divided by the  $^{137}$ Cs concentration in the composite effluent. The DF for  $^{137}$ Cs was 7.78 E+5, and is equivalent to an effluent with a  $^{137}$ Cs concentration of 1.1E-4  $\mu$ Ci/mL. The maximum DF, 1.4 E+6, measured the best performance that could be expected from this column system.

The LAW vitrified waste form must be no greater than 0.3 Ci  $^{137}$ Cs/m $^3$ . This limit can be converted to  $^{137}$ Cs concentration in the Cs ion exchange effluent of 0.125  $\mu$ Ci  $^{137}$ Cs/mL. $^3$  The product effluent  $^{137}$ Cs concentration was a factor of 1000 below this contract limit.

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 $<sup>^3</sup>$  The conversion requires the following assumptions: Envelope A LAW will contain 14 wt% Na<sub>2</sub>O, all Na comes from the tank waste, the glass density is 2.66 g/mL, and the waste Na concentration is 5 M Na. For the maximum 20 wt% waste Na<sub>2</sub>O loading, the Cs ion exchange effluent must be less than 0.087  $\mu$ Ci  $^{137}$ Cs/mL.

The lead column was eluted with 0.5 M HNO $_3$  to  $C/C_o$  of 1% in 12 BVs with >94% of the  $^{137}Cs$  contained in 2.5 BVs of eluant. The peak  $^{137}Cs$   $C/C_o$  value was 64 (based on 1-BV collection increments of nominally 10-mL).

### **Quality Requirements**

This work was designated as QL-3 per the RPP-WTP Quality Assurance Program, BNFL-5193-QAP-01, Rev. 6. PNWD implemented the RPP-WTP quality requirements by performing work in accordance with the quality assurance plan, CHG-QAPjP, Rev. 0.

#### **Issue**

U and Pu were found in the Cs eluate, indicating these materials exchanged onto the resin. No additional issues were observed.

# **Terms and Abbreviations**

AV apparatus volume BV bed volume

C/C<sub>o</sub> analyte concentration in column effluent divided by analyte

concentration in feed

CMC Chemical Measurement Center

DI deionized

EQL estimated quantitation level

F-factor mass of dry ion exchanger divided by mass of wet exchanger

FMI Fluid Metering, Inc., Syosset, NY

GEA gamma energy analysis HLRF High Level Radiation Facility

IBC IBC Advanced Technologies, Inc., American Fork, Utah

IC ion chromatography

ICP-AES inductively coupled plasma/atomic emission spectrometry

ICP-MS inductively coupled plasma/mass spectrometry

λ column distribution ratio
 L/D length over diameter ratio
 M molarity, moles/liter
 MDL method detection limit
 MRQ minimum reportable quantity
 NMRQ no minimum reportable quantity

NPT National Pipe Thread PSD particle size distribution

SAL Shielded Analytical Laboratory

TIC total inorganic carbon TOC total organic carbon



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# 1.0 Introduction

The U. S. Department of Energy plans to vitrify tank wastes at the Hanford Site in preparation for permanent disposal. Before vitrification, tank wastes will be divided into low-activity and high-activity fractions through specific pretreatment processes. The pretreatment flow sheet for the Hanford high-level tank wastes includes the use of SuperLig<sup>®</sup> 644 (SL-644) material for <sup>137</sup>Cs removal from the aqueous waste fraction. The SL-644 is a Cs-selective, organic, ion exchanger and has been shown to be effective in removing Cs from a variety of Hanford tank wastes (Hassan, McCabe, and King 2000; Hassan, McCabe, King and Crowder, 2000; Hassan, King, McCabe, and Crowder 2001; King, Hassan, and McCabe 2001; Kurath, Blanchard, and Bontha 2000a; Kurath, Blanchard, and Bontha 2000b). The SL-644 has been developed and supplied by IBC Advanced Technologies, Inc., American Fork, UT.

This report summarizes batch-contact studies of SL-644 and crystalline silicotitanate<sup>4</sup> (IE-911, UOP LLC, Des Plaines, IL), and dual small-column testing of the SL-644 ion exchange material. The test matrix was 241-AP-101 Hanford tank waste diluted to 5 M Na<sup>+</sup> (hereafter referred to as AP-101DF) (Goheen, et al., 2001). Approximately 1.2 L of AP-101DF were processed through the ion exchange column system. The ion exchange process steps tested include resin bed preparation, loading, feed displacement, water rinse, elution, and resin regeneration.

The objectives of this work were to:

- determine batch distribution coefficients as a function of Na/Cs molar concentration for SL-644 and IE-911 in the AP-101DF matrix.
- demonstrate the <sup>137</sup>Cs decontamination of Envelope A tank waste sample AP-101DF using SL-644 columns and provide a Cs-decontaminated sample for downstream process testing (i.e., <sup>99</sup>Tc removal, low activity waste (LAW) melter feed testing, and LAW vitrification).
- develop loading and elution breakthrough profiles for SL-644 columns.
- produce and characterize the Cs eluate solution from SL-644 columns for use in evaporation tests and high-level waste (HLW) vitrification.
- demonstrate the effectiveness of all SL-644 ion exchange process steps including loading, feed displacement, deionized (DI) water washing, elution and resin regeneration.<sup>5</sup>

<sup>4</sup>IE-911 is an alternative Cs ion exchanger and is being tested in parallel to SL-644 only with batch contacts. The IE-911 was developed and supplied by UOP LLC, Des Plaines, IL.

<sup>&</sup>lt;sup>5</sup> Test Specification: "Tank 241-AN-102 and 241-AP-101 Ion Exchange," TSP-W375-00-00028, Rev. 1, M. E. Johnson, CH2M Hill Hanford Group, Dec. 11, 2000 and Test Plan "Actual Waste Ion Exchange Testing for the RPP-WTP Project," CHG-TP-41500-013, Rev. 0, D. L. Blanchard, Jan. 24, 2000.

# 2.0 Experimental

## 2.1 SL-644 Resin

Initial AP-101DF batch contacts were conducted with SL-644 resin batch 981112YK-N3-16/18. This material batch had a distinct salt and pepper appearance with a significant light gray fraction. Initial batch distribution studies with this SL-644 batch on AW-101 simulant indicated it performed well and it was used for AP-101DF batch distribution testing. Unfortunately, later simulant column testing proved this material inadequate (Fiskum, Blanchard, and Arm 2002). Additional SL-644 was obtained from IBC production batch number 010319SMC-IV-73. This material batch exhibited a black-red appearance peppered with light brown specs.

Before sampling SL-644 from the IBC-supplied plastic bottle, the bulk resin bottle was turned several times to produce a well-mixed material. Sub-samples were removed immediately after mixing. The sieve results of the two resin batches were previously reported (Fiskum, Blanchard, and Arm 2002) and are reproduced in Table 2.1. The 212- to 425-μm particle size dry-sieved fraction was used in the ion exchange columns. The average particle size corresponded to 540-μm in diameter, expanded in 3 M NaOH-2 M NaNO<sub>3</sub>-0.1 M KNO<sub>3</sub> (Fiskum, Blanchard, and Arm 2002). As a general rule, the column diameter should be 20 times greater than the resin particle diameter to minimize wall effects (Korkisch). Given the diameter of the column at 1.46 cm, the column diameter was 27 times the average diameter of the 212- to 425-μm diameter dry-sieved resin particles expanded in caustic solution.

<b>Table 2.1.</b> Dry Particle-Size	Weight-Percent Distribution of Various Batches of
	As-Received SL-644

	Particle Size	981112YK-N3-16/18	010319SMC-IV-73
Sieve Size <sup>(1)</sup>	(µm)	wt %	wt %
18	>1000	1.78	0.06
30	600-1000	55.25	37.27
40	425-600	20.12	38.23
50	300-425	13.76	18.01
70	212-300	7.93	6.08
100	150-212	1.11	0.26
140	106 -150	0.02	0.06
>140	<106	0.03	0.03
(1) U. S. standar	d sieve size corresp	onds to ASTM E-11 specification	ation.

Properties of the 981112YK-N3-16/18 and 212- to 425-μm 010319SMC-IV-73 SL-644 resins have been previously reported (Fiskum, Blanchard, Arm 2002) and selected properties are reproduced in Table 2.2. The F-factor is the ratio of the dry mass of exchanger to the initial mass of the exchanger, and was

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<sup>&</sup>lt;sup>6</sup> This particle size distribution was used successfully in AW-101 simulant testing.

determined at the same time the batch contact samples and column resin fractions were weighed. The F-factor was obtained by drying approximately 0.5 g resin, under vacuum, at 50°C to constant mass. The F-factor was determined on the H-form (hydrogen form or acid form) and the as-received form of the resin. The F-factor for the Na-form resin was performed differently because of stability problems observed in prior tests on this resin form (Steimke et al. 2001). Drying to constant mass under vacuum at ambient temperature was considered adequate for removing water from the Na-form resin. The L-factor represents the fractional mass remaining after washing the as-received resin form with 0.5 M HNO<sub>3</sub> and DI water, correcting for residual water content as described above. The I<sub>Na</sub> represents the fractional mass gain upon conversion from the H-form to the Na-form, correcting for water content as described above.

**Table 2.2.** SL-644 Properties

Property	981112YK-N3-16/18	010319SMC-IV-73
Sieve fraction	As-received	212- to 425-μm
Bulk density, g/mL <sup>(1)</sup>	0.80	0.74
F-factor, as-received	0.891	0.8773
L-factor, conversion to H-form, fractional		
mass remaining	0.493	0.538
F-factor, H-form	0.939	0.932
I <sub>Na</sub> , fractional mass gain from H-form to		
Na-form	1.29	1.25

# 2.2 Crystalline Silicotitanate

Crystalline silicotitanate IE-911, Lot 2081000009, was obtained from UOP LLC, Des Plaines, IL. The UOP particle size specification of this material was 30-60 mesh. The material was provided in the hydrogen form. A 5-g aliquot was converted to the sodium form by soaking in 40 mL 1 M NaOH for two days. The NaOH solution was then decanted and the IE-911 was rinsed with five successive volumes of DI water. The resin was then dried at 95°C overnight. The F-factor, 0.9630 was determined at the same time the batch contact samples were weighed. This was obtained by drying approximately 0.5 g at 95°C to constant mass.

2.2

<sup>&</sup>lt;sup>7</sup> After initial drying at ambient temperature under vacuum to constant mass, the resin was heated to 50°C. The heated product appeared (visual inspection) to have degraded thus potentially nullifying subsequent mass measurements.

# 2.3 AP-101 Diluted Feed

Samples of tank 241-AP-101 (Envelope A) waste were received in the High Level Radiation Facility (HLRF) in 2000. The homogenization, dilution, subsampling, and analysis of this material have been previously described (Goheen et al. 2002). The total volume of AP-101DF available for Cs ion exchange processing was about 1.2 L. The feed composition is summarized in Table 2.3.

**Table 2.3.** Composition of AP-101DF (Envelope A)

Cations, M						
Na <sup>+</sup>	4.97 E+0					
K <sup>+</sup>	7.61 E-1					
Cs <sup>+</sup>	3.68 E-5 <sup>(1)</sup>					
Uranyl	1.7 E-4					
Mole	e Ratios					
Na/Cs mole ratio	1.4 E+5					
K/Cs mole ratio	2.1 E+4					
	ons, M					
AlO <sub>2</sub> - (2)	2.59 E-1					
Cl	4.1 E-2					
F <sup>-</sup>	< 1.2 E-1 <sup>(3)</sup>					
$CO_3^{2-}$	4.46 E-1					
CrO <sub>4</sub> <sup>-2 (2)</sup>	2.92 E-3					
NO <sub>2</sub> -	7.1 E-1					
NO <sub>3</sub>	1.68 E+0					
OH <sup>-</sup>	1.94 E+0					
PO <sub>4</sub> <sup>-3 (2)</sup>	1.2 E-2					
$SO_4^{-2}$	4.8 E-2					
Oxalate	< 1 E-2					
TOC, g/L	1.64 g/L					
Radionucli	des, (μCi/mL)					
Total alpha	2.6 E-04					
<sup>60</sup> Co	2.5 E-03					
<sup>90</sup> Sr	7.1 E-02					
<sup>99</sup> Tc	3.93 E-2					
<sup>99</sup> Tc as pertechnetate	3.46 E-2					
$^{137}\mathrm{Cs}$	1.26 E+2					
<sup>154</sup> Eu	< 7 E-4					
<sup>239+240</sup> Pu	1.10E-4					
<sup>243+244</sup> Cm	2.12E-6					
Solution density, g/mL	1.256					
(1) Total Carporated by Cohoon at a	1 2002 may be biased low. Reported					

<sup>(1)</sup> Total Cs reported by Goheen et al. 2002 may be biased low. Reported value does not agree with <sup>137</sup>Cs determination by GEA and application of isotopic distribution (<sup>133</sup>Cs 60.6 wt%, <sup>135</sup>Cs 15.4 wt%, <sup>137</sup>Cs 24.0 wt%) where a total Cs concentration of 6.04 μg/mL or 4.5 E-5 M is calculated. (2) Al, Cr, and P determined by ICP-AES. The anionic form is assumed on

the basis of waste chemistry.

(3) F concentration is an upper bound; co-eluting anions positively interfered with peak integration.

# 2.4 Batch Contacts

The batch contacts were performed with the H-form of SL-644 batch 981112YK-N3-16/18. The Naform of IE-911 was tested in parallel with the SL-644. Initial batch contacts were performed using AP-101DF at three different Cs concentrations. Aliquots of the AP-101DF subsamples were tested without spiking, additional aliquots were spiked with 0.1 M CsNO<sub>3</sub> to obtain stock solutions of nominally 1E-3 M and 5 E-3 M Cs. The initial Cs concentrations in the AP-101DF stock contact solutions and the corresponding Na/Cs and K/Cs mole ratios are given in Table 2.4. After discerning the SL-644 batch 981112YK-N3-16/18 was insufficient to remove Cs under column loading conditions (Fiskum, Blanchard, and Arm 2002), another batch contact with AP-101DF was performed using the H-form of SL-644 batch 010319SMC-IV-73, 212- to 425-µm dry particle size. The scope of this second test was much reduced in order to conserve AP-101DF feed for use in subsequent column ion exchange and vitrification tests. Thus only the mid-range Cs concentration was tested in duplicate.

**Table 2.4.** Initial Cs Concentrations in the AP-101DF Solutions Used for the Batch Distribution Tests

Solution	Target Initial Cs Conc. [M]	Target Nominal Na/Cs <sup>(1)</sup> mole ratio	Target Nominal K/Cs <sup>(1)</sup> mole ratio					
Solution		12YK-N3-16/18 and IE-9						
Un-spiked	3.68E-5	1.4 E+5	2.1 E+4					
Cs Spike 1	1.0 E-3	5.0 E+3	7.6 E+2					
Cs Spike 2	5.0 E-3	1.0 E+3	1.5 E+2					
•	SL-644 Batch 010319SMC-IV-73							
Cs Spike 3	1.0 E-3	5.0 E+3	7.6 E+2					
(1) Na <sup>+</sup> and K <sup>+</sup> are the primary cations that compete with Cs <sup>+</sup> for ion exchange with SL-644 and								
IF-911		-	-					

The batch-contact tests were performed in duplicate at a phase ratio of approximately  $100 \, \text{mL/g}$  (liquid volume to exchanger mass). Typically,  $0.1 \, \text{g}$  of SL-644 exchanger was contacted with  $10 \, \text{mL}$  of AP-101DF, and  $0.05 \, \text{g}$  of IE-911 and SL-644 for Cs Spike 3 were contacted with  $5 \, \text{mL}$  AP-101DF. The exchanger mass was determined to an accuracy of  $0.0002 \, \text{g}$ . The waste volume was transferred by pipet and the actual volume was determined by mass difference with an accuracy of  $0.0002 \, \text{g}$  and the solution density. Samples were agitated in an orbital shaker for approximately 24 hours for SL-644 and 72 hours for IE-911. The temperature was not controlled but was generally constant at  $26 \pm 1 \, ^{\circ}\text{C}$  during the 3 days of contact.

All batch-distribution measurements were determined by measuring <sup>137</sup>Cs on both the stock solution (initial concentration) and the contacted solution (final concentration). The initial Na and K concentrations were determined with inductively coupled plasma-atomic emission spectrometry (ICP-AES). Initial <sup>133</sup>Cs concentrations were confirmed by inductively coupled plasma-mass spectrometry (ICP-MS).

The batch distribution coefficient, K<sub>d</sub> (with units of mL/g), was determined using the following relationship:

$$K_d = \frac{(C_0 - C_1)}{C_1} * \frac{V}{M * F}$$
 (1)

 $C_0$  = initial <sup>137</sup>Cs concentration  $C_1$  = final <sup>137</sup>Cs concentration

V = volume of the liquid sample (mL)

M = mass of the ion exchanger (g) (SL-644 H-form mass corrected for the Na-form mass

increase or Na-form IE-911 mass)

= dried resin mass divided by the initial resin mass.

The dry bed resin density,  $\rho$ , was obtained according to the following equation:

$$\rho = \frac{m * L * F * I_{Na}}{BV} \tag{2}$$

 $\rho$  = dry Na-form g of resin per mL of resin where

m = resin mass in the column, as-received form

L = fractional mass remaining after washing (0.538)

F = water loss factor, as-received form (0.877)

 $I_{Na}$  = fractional mass gain on conversion from H-form to Na-form (1.25) (this factor is set to 1

when calculating the dry bed density in the H-form or 0.5 M HNO<sub>3</sub> feed)

BV = resin bed volume in the feed (discussed in section 3.2.7).

The Cs  $\lambda$  value (column distribution ratio) is a function of the feed condition batch distribution coefficient and the dry bed density, and is obtained as shown in Equation 3.

$$\lambda = K_d * \rho \tag{3}$$

#### 2.5 Column Run Experimental Conditions

#### 2.5.1 Ion Exchange Column System

A schematic of the ion exchange column system is shown in Figure 2.1. The system consisted of two small columns containing the ion exchange material, a small metering pump, three valves, a pressure gauge, and a pressure relief valve. Valves 1, 2, and 3 were three-way valves that could be turned to the flow position, sample position, or no-flow position. Valve 1 was placed at the outlet of the pump and was used to eliminate air from the system, purge the initial volume of the system, or isolate the columns from the pump. Valves 2 and 3 were primarily used for obtaining samples and could also be used to isolate the

columns from the rest of the system. The columns were connected in series with the first column referred to as the lead column and the second column referred to as the lag column.

The columns were prepared at Savannah River Technology Center Glassblowing Laboratory. Each column consisted of a 15-cm glass column with a 24/40 taper ground-glass fitting on top and a threaded fitting on the bottom. A polyethylene bushing was installed in each glass-threaded fitting to accommodate a 1/4-in stainless steel National Pipe Thread (NPT) fitting. The inside diameter of each column was 1.46 cm, which corresponded to a volume of 1.67 mL/cm. A stainless steel, 200-mesh screen supported the resin bed. A decal millimeter scale affixed to the column allowed for measurement of resin bed height and thus shrinkage and swelling. The upper section contained four entry ports and a taper joint with screw cap that securely fitted the column. The lead column assembly used a pressure relief valve (10 psi trigger), pressure gauge, and sample inlet; the remaining port was plugged. The lag column assembly used one port for sample entry, and the other three ports were plugged. In both columns, the inlet sample lines extended through the port opening to the top of the column. The connecting tubing was 1/8-in OD, 1/16-in ID polyethylene. Valved quick-disconnects (Cole Parmer, Vernon Hills, IL) were installed in-line to allow for ease of column switching. A FMI QVG50 pump (Fluid Metering, Inc., Syosset, New York) equipped with a ceramic and Kynar<sup>®</sup> coated low-flow piston pump head was used for all fluid introduction. The flow rate was controlled with a remotely operated FMI stroke-rate controller. The pump was calibrated with the stroke-rate controller and could provide pumping rates from 0.08 to 16 mL/min. The volume actually pumped was determined using the mass of the fluid and the fluid density. The pressure indicated on the pressure gauge remained below 5 psi during all runs. The total holdup volume of the Cs ion exchange system was the summed volume of all fluid-filled parts and was estimated to be 42 mL.

# 2.5.2 Resin Bed Preparation

The initial SL-644 resin and bed preparations were performed using a procedure similar to that recommended by Hassan et al. (1999). Before packing the resin beds, aliquots of SL-644 resin were placed in separate beakers. The two resin aliquots (4.17 and 4.20 g lead and lag columns, respectively) were soaked in about 40 mL of 1 M NaOH (~10:1 solution to resin volume ratio) with light agitation for 2 h. The NaOH was decanted, and the SL-644 was slurried with about 40 mL of DI water. The resin was then quantitatively transferred into the columns using additional DI water to aid the transfer. Both of the resin beds were individually washed with 0.5 M HNO<sub>3</sub>. After resin washing, the corrected sodium form SL-644 mass was calculated to be 2.46 g and 2.48 g (dried at ambient temperature under vacuum) in the lead and lag columns, respectively, according to the following equation:

$$M_{bed} = M_{AR} * F_{AR} * L * I_{Na}$$
 (4)

where  $M_{bed}$  = resin bed mass

 $M_{AR}$  = resin mass loaded in the column, as-received form, 4.17 g lead and 4.20 g lag columns

 $F_{AR}$  = F-factor for the as-received resin, 0.877

L = fractional mass remaining after conversion to the H-form, 0.538

 $I_{Na}$  = fractional mass gain on conversion from the H-form to the Na-form, 1.25.

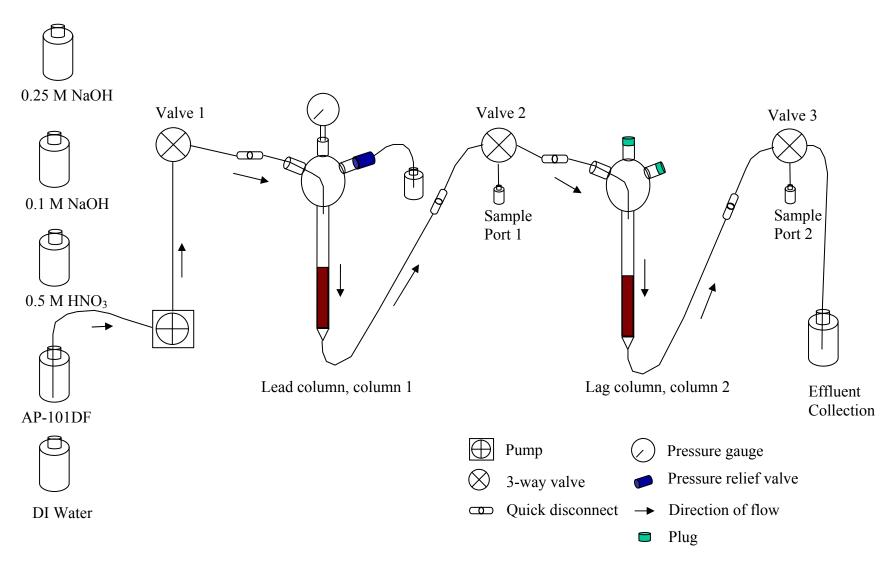


Figure 2.1. Cesium Ion Exchange Column System

The entire ion exchange system was rinsed with DI water, regenerated with 0.25 M NaOH, and then used for a full shakedown experiment with AW-101 simulant (Fiskum, Blanchard, and Arm 2002). Both columns were individually eluted and rinsed. A summary of the shakedown test operations is provided in Table 2.5.

**Table 2.5.** Bed Conditioning and Simulant Shakedown Testing Parameters

			rage Volu	Flow Rate,				
Process Step	Solution	BV <sup>(1)</sup>	$AV^{(2)}$	mL	BV/h			
Pre-Conditioning, Lead and Lag Columns Separate								
Wash	0.5 M HNO <sub>3</sub>	13.5		150	0.87			
DI water rinse	DI water		3.0	67	2.1			
Regeneration	0.25 M NaOH	5.1		56	2.0			
Loading, Feed Displac	ement, and Rinse, Co	lumns in S	eries					
Regeneration (cont.)	0.25 M NaOH	3.7		41	1.9			
Load	AW-101 simulant	143		1576	2.6			
Feed Displacement	0.1 M NaOH		2.9	121	2.5			
Rinse	DI water		2.0	84	2.5			
Elution, Lead and Lag Columns Separate								
Elution	0.5 M HNO <sub>3</sub>	12		132	0.92			
Eluant rinse	DI water		2.7	60	2.6			

<sup>(1)</sup> BV = bed volume (11 mL in 0.25 M NaOH regeneration condition)

# 2.5.3 Ion Exchange Column Test with AP-101DF

The beds had been stored approximately 3 weeks in the H-form in DI water since the end of the simulant AW-101 column run. The resin beds were fluidized with DI water. Then the ion exchange system was loaded into the Shielded Analytical Laboratory (SAL) hot cell with the resin in the H-form. All subsequent processing was performed in the hot cells at temperatures ranging from 25 to 29°C. The entire system was initially cycled in the hot cell with 0.25 M NaOH, DI water, 0.5 M HNO<sub>3</sub>, and DI water sequentially through the lead and lag columns. The bed conditioning, AP-101DF loading, 8 feed displacement, and DI water rinse steps were conducted by passing these solutions through both resin beds connected in series. The elution, elution rinse, and the regeneration steps were conducted on the lead column only. The experimental conditions for each process step are shown in Table 2.6. The resin bed volume (BV) is the volume in 0.25 M NaOH. For the AP-101DF test, the BV was measured at 10.6 mL in the regeneration condition just prior to waste loading. The AP-101DF effluent was collected in four effluent bottles. The first bottle collected 64 mL, slightly over one apparatus volume (AV), and consisted primarily of the displaced regeneration solution. The following effluent collections captured nominally

2.9

<sup>(2)</sup> AV = apparatus volume (42 mL for columns in series; 23 mL for lead column processing and 21 mL for lag column processing)

<sup>&</sup>lt;sup>8</sup> AP-101DF processing represents the second Cs load and elute cycle for the resin beds.

350-mL AP-101DF fractions. After ascertaining <sup>137</sup>Cs activity was low, the three 350-mL fractions were combined into one composite for follow-on Tc ion exchange.

The solution above the lead resin bed was drained to within about 1 cm of the resin surface when switching to the next process step. Then nominally 5-mL of the solution were allowed to build up above the resin bed prior to commencement of flow. This helped minimize reagent mixing thus speeding the reagent transition through the column system.

**Table 2.6.** Experimental Conditions for AP-101DF Ion Exchange

		Total Volume		Flow rate		Time,	T,	
Process step	Solution	$BV^{(1)}$	$AV^{(2)}$	mL	BV/h	mL/min	h	°C
Bed conditioning	0.25 M NaOH	9.8	2.5	104	1.9	0.33	5.2	25
Bed conditioning	DI water	5.7	1.4	60.6	2.7	0.48	2.1	25
Bed conditioning	0.5 M HNO <sub>3</sub>	7.8	2	82.2	2.5	0.44	3.1	25
Bed conditioning	DI water	6.9	1.8	74	2.8	0.49	2.5	25
Regeneration	0.25 M NaOH	9.6	2.4	103	1	0.18	9.6	25
Loading lead column	AP-101 Feed	112.2	-	1189	2.8	0.50	37.8	26-28
Loading lag column <sup>(3)</sup>	AP-101 Feed	109.4	-	1159	2.8	0.50	37.8	26-28
Feed displacement	0.1 M NaOH	13.2	3.3	140	2.74	0.48	5.1	28
DI water rinse	DI water	8.7	2.2	92	2.9	0.51	3.2	28
Elution of lead colum	n							
Elution	0.5 M HNO <sub>3</sub>	19.3	-	204	0.88	0.16	24.6	28-29
Eluant rinse	DI water	6.9	3.19	73.7	2.9	0.51	4.5	28-29
Regeneration	0.25 M NaOH	4.7	2.17	50	0.9 - 1.5	0.16 - 0.27	6.2	23
Rinse	DI water	5	2.31	53.1	2.59	0.46	1.9	24

<sup>(1)</sup> BV in 0.25 M NaOH regeneration condition (10.6 mL),

The sampling and analysis protocol is shown in Table 2.7. During the loading phase, small samples (about 2 mL) were collected from the lead column starting at 7 BVs, then at 15 BVs and continuing in nominal 10-BV increments of feed from both the lead and lag columns. The flow rate ranged from 2.99 to 3.23 BV/h in the feed condition, which was equivalent to 2.74 to 2.96 BV/h when the resin bed was expanded in the regeneration solution. Flow rate doubled during sample collection from the lead column because of the strong siphoning effect at the sample port. Thus, for the 13 samples collected from the lead column, and for 2 minutes each sample, AP-101DF fed at a nominal flow rate of 6 BV/h. The feed displacement, DI water rinse, elution, and elution rinse samples were taken at 1-BV increments at flow rates shown in Table 2.6.

<sup>(2)</sup> AV (42 mL for sequential column format, 23 mL for lead column only)

<sup>(3)</sup> The feed volume through the lag column is reduced because of sampling from the lead column.

Table 2.7. Sampling Interval and Analyses

			Approximate	
<b>Process Step</b>	Lead Column	Lag Column	Sample Size (mL)	Analyses
Bed conditioning	_	_	ı	_
Bed conditioning	_	_	1	_
Bed conditioning	_	_	1	_
Loading	Every 10 BVs	Every 10 BVs	2	GEA
Feed displacement	_	Every 1 BVs	10	GEA
DI water rinse	_	Every 1 BVs	10	GEA
Elution	Every 1 BVs	_	10	GEA
Eluant rinse	Every 1 BVs	_	10	GEA
Regeneration	1 composite	_	50	ICP-AES, GEA, OH-
		<b>Composite Sar</b>	nples	
Effluent – 1	_	_	2	GEA
Effluent – 2	_	_	2	GEA
Effluent – 3	_	_	2	GEA
Effluent – 4	_	_	2	GEA
				GEA, ICP-AES, IC,
Effluent composite				U, ICP-MS (Tc),
(effluents 2 - 4)	_	_	10	pertechnetate
				ICP-AES, GEA,
				TOC, IC, U,
				ICP-MS, and various
Eluate	1 composite	-	10	radioisotopes

# 2.5.4 Sample Analysis

The <sup>137</sup>Cs concentration was determined using a bench-top GEA spectrometer. This allowed for rapid sample analysis. Selected sample results were later confirmed by GEA analysis in the analytical laboratory. Because of the high dose rate from <sup>137</sup>Cs, the feed sample and eluate samples required dilution before removal from the hot cell. These samples were diluted with 0.25 M NaOH and 0.5 M HNO<sub>3</sub>, respectively. The extent of dilution was determined by mass difference. Composite samples from each of the effluent bottles were taken for analysis by GEA. Another composite sample was taken for analysis once the contents of effluent bottles 2 through 4 were combined for Tc ion exchange. The regeneration solution was composited and a sample taken for ICP-AES and OH<sup>-</sup> determination. Once the GEA results were confirmed for the eluate samples, the eluate samples were composited and a sample of the composite was submitted for various analyses; GEA, ICP-AES, total organic carbon (TOC), ion chromatography (IC), U, pertechnetate, and ICP-MS for <sup>99</sup>Tc.

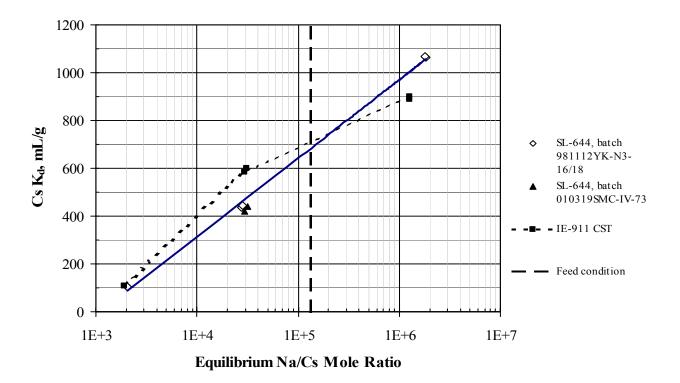
The Na and other metal concentrations were determined with ICP-AES. The OH concentration was determined by potentiometric titration with standardized HCl. Uranium was determined using kinetic phosphorescence. The total Tc concentration was determined by ICP-MS and the pertechnetate concentration was determined using radiochemical separations specific for pertechnetate followed by beta counting. Anions were determined using IC. TOC and total inorganic carbon (TIC) were determined by silver-catalyzed hot persulfate oxidation and furnace oxidation methods.

# 3.0 Results and Discussion

### 3.1 Batch Contact Results

The  $K_d$  values were calculated using Equation 1 and are based on the  $^{137}$ Cs concentrations as measured by GEA. The Na/Cs mole ratios were calculated based on the measured Na and total Cs concentrations in the uncontacted AP-101DF. The equilibrium Cs concentration was based on the measured  $^{137}$ Cs concentrations and the ratio of  $^{137}$ Cs:total Cs determined for the unspiked and spiked solutions. The Na concentration was assumed to be constant for the batch contacts. Since the quantity of  $H^+$  added with the resin was small relative to the moles of  $Na^+$  and  $OH^-$  in the contact solution (phase ratio of 100 mL of solution:gram of exchanger) this is a reasonable assumption. In these experiments the waste solutions were estimated to have 19.4 meq of  $OH^-$  and  $OH^-$  and  $OH^-$ , while the resin contained  $OH^-$  and  $OH^-$  for the  $OH^-$  for the  $OH^-$  and  $OH^-$  for the  $OH^-$  for the

The calculated  $^{137}$ Cs  $K_d$  values are plotted in Figure 3.1 for SL-644 and IE-911 in contact with AP-101DF. The IE-911 feed condition  $K_d$  value is interpolated at 710 mL/g. The SL-644, 981112YK-N3-16/18 resin batch  $K_d$  values are defined by the following equation (linear least squares fit):  $K_d = 142.44 * Ln([Na]/[Cs]) - 997.49$ ,  $R^2 = 0.998$ .



**Figure 3.1.** <sup>137</sup>Cs Distribution Coefficients ( $K_d$ ) for AP-101DF ( $T = 26^{\circ}C$ )

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<sup>&</sup>lt;sup>9</sup> The H-form resin contains 2.2 meg H<sup>+</sup> per gram. (Rapko, et al. 2002)

The experimental data from the single Cs concentration tested (Cs spike 3) shows the  $K_d$  value, represented by the 010319SMC-IV-73 SL-644 resin batch (212- to 425- $\mu$ m particle size), is equivalent to the 981112YK-N3-16/18 material. However the slope could not be evaluated. Assuming equivalent slope, the  $K_d$  value at the AP-101DF feed condition can be estimated at 690 mL/g from the point at which the  $K_d$  curve crosses the Na/Cs mole ratio (1.4 E+5) for the AP-101DF sample.

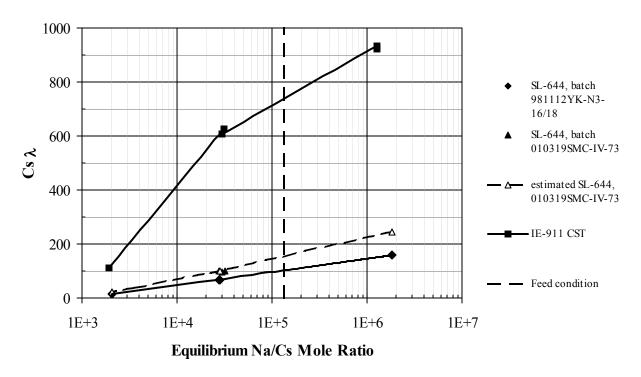
The calculated dry-bed densities determined using the resin masses and volumes in the ion exchange column system are given in Table 3.1. Good agreement was obtained between the lead and lag columns. The 0.25 M NaOH dry-bed density was equivalent to the 0.5 M HNO<sub>3</sub> form dry-bed density. In this case, the decrease in mass associated with the change from the Na-form to the H-form was equivalent to the decrease in volume upon contracting from the Na-form to the H-form.

**Table 3.1.** Dry Bed Density

	010319SMC-IV-73,	010319SMC-IV-73,
Property	Lead Column	Lag Column
Column resin mass, <sup>(1)</sup> g	4.17	4.20
Corrected column Na-form resin mass, g	2.46	2.48
Bed volume, 0.25 M NaOH, mL	10.7	10.5
Bed volume, AP-101DF, mL	9.7	9.7
Bed volume, 0.5 M HNO <sub>3</sub> , mL	7.7	not performed
Dry bed density, ρ, in given matrix		
Na-form resin		
0.25 M NaOH, g/mL	0.230	0.236
AP-101DF, g/mL	0.254	0.255
H-form resin		
0.5 M HNO <sub>3</sub> , g/mL	0.256	not performed
(1) The 212- to 425-µm particle-size resin mass me	easured in the as-received form.	

The calculated Cs  $\lambda$  values, as a function of Na/Cs mole ratio, are shown in Figure 3.2. Two Cs  $\lambda$  curves are plotted for SL-644. One curve represents the measured K<sub>d</sub> values from SL-644 batch 981112YK-N3-16/18 multiplied by the calculated resin bed density of 0.15 g/mL (Fiskum, Blanchard, and Arm 2002). The other SL-644 Cs  $\lambda$  curve is estimated for the 010319SMC-IV-73 resin, applying the resin bed density of 0.23 g/mL to the K<sub>d</sub> values generated for the 981112YK-N3-16/18 resin. The Cs  $\lambda$  value is estimated to be 170 BVs in the AP-101DF feed condition, and 160 BVs in the 0.25 M NaOH regeneration condition for 010319SMC-IV-73 SL-644. This value is the approximate point at which the Cs breakthrough curve was predicted to pass through C/C<sub>o</sub> = 50% if sufficient feed was available for processing.

The calculated Cs  $\lambda$  values for IE-911 are also shown in Figure 3.2. The IE-911 bed density was determined to be 1.04 g/mL. The feed condition Cs  $\lambda$  value for IE-911 is estimated to be 740 BVs.



**Figure 3.2.** Estimated <sup>137</sup>Cs  $\lambda$  Values for AP-101DF (T = 26°C)

# 3.2 Column Test

#### 3.2.1 Initial Resin Conditioning

The resin bed conversion progress from Na-form to H-form and vice versa could be monitored visually according to the color of the resin bed. The Na-form appeared black in the columns whereas the H-form appeared brown. Initial conversion to the Na-form in the sequential column format required nominally 104 mL (9.8 BVs fully expanded) or 2.5 AVs. During this time the BVs increased from 7.5 mL to 11 mL. Conversion to the H-form required 82 mL of 0.5 M HNO<sub>3</sub> (7.8 BV fully expanded) or 2.0 AVs. During this time the resin bed shrank from 11 mL to 7.8 mL. The second regeneration was performed at half the flow rate (1 BV/h) of the first regeneration. The same volume of 0.25 M NaOH was required to complete the conversion, as evidenced by visual examination.

#### 3.2.2 Loading

After the resin beds were conditioned, the loading phase was initiated with AP-101DF. Approximately 1.5 AVs (64 mL) of effluent were initially collected in a separate collection bottle. Most of this effluent was 0.25 M NaOH from the regeneration step mixed with some AP-101DF in the ion exchange apparatus. This solution was maintained separately and not mixed with the final Cs-decontaminated AP-101DF effluent composite; thus most of the apparatus 0.25 M NaOH solution was

prevented from mixing with the AP-101DF effluent. The resin shrank about 8.5% to 9.7 mL in the AP-101DF feed.

The Cs effluent concentrations from the columns are shown in Figure 3.3 as %  $C/C_o$  vs. the BVs of feed processed through each column. The abscissa reflects BVs as a function of the resin in the expanded regeneration condition of 10.6 mL. The  $C_o$  value for  $^{137}$ Cs was determined to be 126  $\mu$ Ci/mL. The  $C/C_o$  is plotted on a probability scale as this scale tends to provide a straight-line breakthrough curve. The  $C/C_o$  values, determined using a bench-top GEA spectrometer, were generally in good agreement with selected samples independently analyzed by the analytical laboratory. Raw analytical results and calculations are located in the appendix.

The contract  $^{137}$ Cs removal limit is also shown in Figure 3.3. The C/C<sub>o</sub> value of 0.099% corresponds to the contract limit of 0.3 Ci/m<sup>3</sup> for  $^{137}$ Cs in the LAW glass. The C/C<sub>o</sub> value corresponding to this limit is determined using the Na concentration of 5.0 M in the AP-101DF, a  $^{137}$ Cs feed concentration of 126  $\mu$ Ci/mL, a 14 wt% waste Na<sub>2</sub>O loading in the glass, and a glass product density of 2.66 g/mL.  $^{10}$ 

The 50% Cs breakthrough value is the point at which the  $C/C_o$  is 50% (0.5) and is normally a direct indicator of the effective capacity of the resin. A straight-line extrapolation to 50% breakthrough from 0.2%  $C/C_o$  is inherently prone to error. With this caveat, the 50% Cs breakthrough extrapolated to about 190 BVs (0.25 M NaOH condition) for the lead column. This value is in good agreement with the predicted Cs  $\lambda$  value of 160 BVs determined with batch-contact tests. There is insufficient data for an estimate of the 50% Cs breakthrough for the lag column.

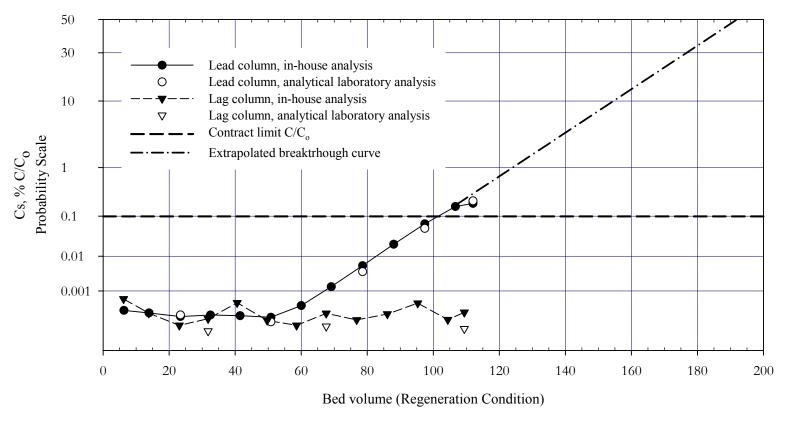
The decontamination factors (DFs) were calculated on composites from each of the three effluent collection bottles and are summarized in Table 3.2. These may be compared to the contractual limit of  $C/C_o = 0.099\%$  (DF= 1,010). In all cases, sufficient Cs was removed that the contract limit for Cs removal was met. The final lag column sample  $C/C_o$  and DF are also shown. The DF result is equivalent to the maximum DF within experimental error.

Table 3.2.	Decontamination	Factors for 13	Cs from	AP-101DF
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Sample	Volume, BV (mL)	<sup>137</sup> Cs Concentration μCi/mL <sup>(1)</sup>	C/C <sub>0</sub> , %	DF
Final lag column sample	109 (1160)	7.4 E-05	7.3 E-05	1.4 E+06
Effluent composite bottle-2	34 (356)	9.8 E-05	7.7 E-05	1.3 E+06
Effluent composite bottle-3	36 (377)	1.1 E-04	8.6 E-05	1.2 E+06
Effluent composite bottle-4	40 (424)	2.3 E-04	1.8 E-04	5.5 E+05
Composite effluent	104 (1103)	1.08 E-04	8.6 E-05	1.2 E +06
(1) The <sup>137</sup> Cs uncertainty ranges from	om 3% to 8% relativ	e error, 1-σ.		

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 $<sup>^{10}</sup>$  Applying a glass loading of 20 wt% would reduce the contract limit % C/C $_{\rm o}$  to 0.069 %.



Conditions: SL-644 batch number 010319SMC-IV-73 212- to 425- $\mu$ m dry particle size Process temperature = 28 °C

BV in 0.25 M NaOH feed condition = 10.6 mL

Flow Rate = 2.8 BV/h

 $^{137}$ Cs  $C_o = 126 \mu Ci/mL$ 

Na concentration = 4.97 M.

**Figure 3.3.** <sup>137</sup>Cs Breakthrough Curves for AP-101DF Sample, Probability Plot

The three composite effluent fractions were combined into a single composite for subsequent processing. This Cs-decontaminated effluent was characterized and the results are summarized in Table 3.3. Major constituents (Na, K, Al, NO<sub>3</sub>, and NO<sub>2</sub>) in the effluent were similar in concentration to the feed within experimental error. The authors have no explanation for the  $\sim$ 25% increase measured for total Tc and OH other than experimental/analytical error. The pertechnetate remained constant within experimental error. Thus the total Tc determined by ICP-MS may be biased high in this analysis and/or biased low in the feed analysis. Notably nearly all U and Pu<sup>12</sup> were removed from the AP-101DF by SL-644. The F, PO<sub>4</sub> and SO<sub>4</sub> also resulted in low effluent recovery. The low recovery is thought to be due to analytical error/uncertainty.

# 3.2.3 Feed Displacement and Rinse

There was a delay between the introduction of the feed displacement and the collection of these samples from the effluent line. Approximately 42 ml of Cs-decontaminated AP-101DF were collected as effluent prior to starting the feed displacement sampling. The relative Cs concentration, as  $C/C_o$ , was < 4 E-3 % for all feed displacement and DI water rinse samples.

#### 3.2.4 Elution and Eluant Rinse

The elution proceeded similarly as found in the simulant test with respect to the Cs elution profile and resin bed shrinkage. However, the characteristic color change from black in the Na-form to tan in the H-form was not visible. The resin appeared virtually black for the entire elution and rinse cycles.

The lead column  $C/C_o^{13}$  values for  $^{137}Cs$  are shown in Figure 3.4 for the elution and the eluant rinse steps. The ordinate is a logarithmic scale to clearly show the large range of  $C/C_o$  values obtained. The abscissa is given in BVs relative to the regeneration condition. The majority of the  $^{137}Cs$  was contained in elution BVs 3-5. The peak value of  $C/C_o$  was found to be 64. The elution cutoff of  $C/C_o = 0.01$  was reached at 12 BVs but elution was continued beyond this due to the lag between sample collection and the determination of the Cs concentration. At 18.5 BVs the  $C/C_o$  visibly rises. At this point, the columns were allowed to sit in 0.5 M HNO<sub>3</sub> for about two hours while samples were counted. Cesium continued to elute from the resin and when flow was re-started, showed a slight rise in relative Cs concentration. The  $C/C_o$  values for the eluant rinse with DI water dropped rapidly in  $^{137}Cs$  concentration indicating DI water did not continue Cs elution.

 $<sup>^{11}</sup>$  Fiskum et al. 2000 reported undiluted AP-101 tank waste total  $^{99}Tc$  concentration of 4.90E-2  $\mu Ci/mL$ . Applying the dilution factor of 1.13 to this value results in a calculated  $^{99}Tc$  concentration of 4.34E-2  $\mu Ci/mL$  in the AP-101DF. The measured feed value of 3.93E-2  $\mu Ci/mL$  is 10% lower than the calculated value, indicating a possible slight low bias in the measured feed concentration.

possible slight low bias in the measured feed concentration.

12 Pu concentration in the effluent was estimated from the Pu concentration in the Tc ion exchange effluent. This assumes that the Tc ion exchange processing did not remove Pu.

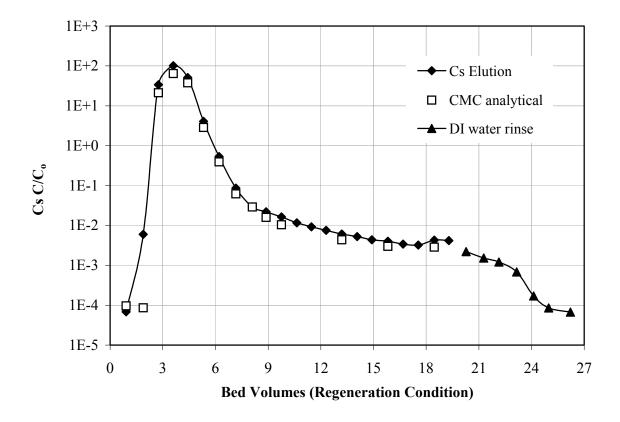
assumes that the Tc ion exchange processing did not remove Pu.  $^{13}$  The  $C_o$  refers to the  $^{137}$ Cs concentration in the AP-101 DF sample fed to the lead column. For elution, the  $C/C_o$  value is an indication of the extent to which  $^{137}$ Cs is concentrated relative to the feed. It is an indirect measure of the extent to which the resin is actually eluted.

Table 3.3. AP-101DF Cs-Decontaminated Product Effluent

Component	Concentration, M	% Change from Feed
Cations		
Na <sup>+</sup>	4.81 E+0	-3.0
K <sup>+</sup>	7.44 E-1	-2.2
Uranyl	3.8 E-6	-98
Anions		•
AlO <sub>2</sub> -	2.45 E-1	-5.3
F <sup>-</sup>	9.74 E-2	-19.
Cl	4.09 E-2	-1.2
CrO <sub>4</sub> <sup>-2</sup>	2.70 E-3	-7.5
$NO_2^-$	7.78 E-1	10.
$NO_3$	1.88 E+0	12.
OH-	2.40 E+0	24
$PO_4^{-3}$ (IC)	7.3 E-3	-27.
P (ICP-AES)	1.06 E-2	-15.
$SO_4^{-2}$	3.25 E-2	-33.
Oxalate	1.1 E-2	-2.7
Radionuclides	μCi/mL	% Change from Feed
<sup>60</sup> Co	2.23 E-3	-12.
<sup>99</sup> Tc	4.99 E-02	27.
<sup>99</sup> Tc as pertechnetate	3.95 E-2	14.
<sup>137</sup> Cs	1.08 E-4	>-99
<sup>154</sup> Eu	< 3.0 E-5	NA
<sup>239+240</sup> Pu	1.74E-5 (estimated) (1)	-84 <sup>(1)</sup>
Solution density, g/mL	1.257	0.08

NA = not applicable, not detected in feed or effluent

<sup>(1)</sup> Pu removal estimated based on feed Pu concentration and effluent Pu concentration from Tc ion exchange. The <sup>239+240</sup>Pu concentration in the Tc-removed AP-101DF effluent was 1.74E-5 μCi/mL. The Pu removal estimate is supported by the eluate analysis where 72% Pu was recovered.



Conditions: SL-644 batch number 010319SMC-IV-73 212- to 425-μm dry particle size

Process temperature = 28-29 °C

BV in 0.25 M NaOH feed condition = 10.6 mL

Flow Rate = 0.88 BV/h

 $^{137}$ Cs C<sub>o</sub> = 126  $\mu$ Ci/mL

Eluant is 0.5 M HNO<sub>3</sub>.

Figure 3.4. <sup>137</sup>Cs Elution and Eluant Rinse of the Lead Column

The eluate samples from the lead column were composited and a sub-sample taken for analysis. The remainder of the composite was saved for vitrification testing. The analytical results are shown in Tables 3.4 and 3.5. Sodium was the dominant component detected using ICP-AES with some Ca, Cd, Cr, Fe, K, Pb, U, and Pu eluting as well. The relative fractions of these analytes recovered in the eluate are given in the tables. The U recovered in the eluate (~42 mg) represents 87% of the U loaded in the AP-101DF feed (48 mg). Plutonium (based on <sup>239+240</sup>Pu) is also largely recovered (72%) in the eluate. A significant fraction (57%) of the Cm was also recovered in the eluate. The other analytes recovered to a much

 $<sup>^{14}</sup>$  The sum of  $^{239}\text{Pu}$  and  $^{240}\text{Pu}$  measured by ICP-MS (8.98E-4  $\mu\text{Ci/mL})$  was a factor of two higher than the radiochemistry determination of the  $^{239+240}\text{Pu}$  (4.63E-4  $\mu\text{Ci/mL})$ . The ICP-MS determination is considered biased high.

smaller percentage in the eluate (Na: 0.14%; Ca: 2%; Cr: 1%; Fe: 29%; K: 0.03%; Pb: Sr: 7%). Because the feed concentrations were "less than" values, the Cd and Pb recovered at >10% and >8%, respectively. The B and Si were probably leached from the glass vials used during the sample preparation process; the preparative blank sample was nearly as high in B and Si concentrations as the eluate sample. Thus the reported B and Si concentrations are not indicative of the eluate composition. As expected <sup>137</sup>Cs was the dominant radionuclide detected. The only anion detected was nitrate, which is not surprising since the eluant was 0.5 M HNO<sub>3</sub>.

Most of the specified minimum reportable quantity (MRQ) levels were met with some exceptions. The large amount of  $^{137}$ Cs prevented the detection limits for  $^{241}$ Am,  $^{154}$ Eu, and  $^{155}$ Eu from meeting the MRQ levels. Relatively high  $^{137}$ Cs activity increased the gamma background level in the detectors due to Compton scattering, thereby making it difficult to detect lower concentrations of other gamma emitters. The large amount of  $NO_3$  prevented the detection limit for Cl from meeting the MRQ level of 3  $\mu$ g/mL. The large  $NO_3$  concentration required large sample dilutions and increased the method detection limit for the other components. The anion concentrations other than nitrate are expected to be small. The TIC analysis was not completed because carbonate is known to evolve as  $CO_2$  in acidic solutions.

**Table 3.4.** Inorganic and Organic Analytes in the Lead Column Eluate Composite

Analyte	Method	MRQ μg/mL	Cs eluate µg/mL	% recovered	Analyte	Method	MRQ μg/mL	Cs eluate µg/mL	% recovered
Ag	ICP-AES	NMRQ	< 0.63	ND	Rh	ICP-MS	NMRQ	5.7 E-2	NM
Al	ICP-AES	75	18.6	0.05	Ru	ICP-MS	NMRQ	3.2 E-3	NM
As	ICP-AES	NMRQ	< 6.3	ND	Sb	ICP-AES	NMRQ	< 13	ND
В	ICP-AES	NMRQ	194	НВ	Se	ICP-AES	NMRQ	< 6.3	ND
Ba	ICP-AES	2.3	[0.29]	[>4] <sup>(1)</sup>	Si	ICP-AES	170	145	НВ
Be	ICP-AES	NMRQ	< 0.25	ND	Sn	ICP-AES	1500	< 38	ND
Bi	ICP-AES	NMRQ	< 2.5	ND	Sr	ICP-AES	NMRQ	< 0.38	ND
Ca	ICP-AES	150	[25]	[2.0]	Та	ICP-MS	NMRQ	2.7E-3	NM
Cd	ICP-AES	8	[1.2]	[>10] <sup>(1)</sup>	Te	ICP-AES	NMRQ	< 38	ND
Ce	ICP-AES	NMRQ	< 5.0	ND	Th	ICP-AES	NMRQ	< 25	ND
Co	ICP-AES	30	< 1.3	ND	Ti	ICP-AES	17	< 0.63	ND
Cr	ICP-AES	15	12	1.4	U	ICP-AES	600	[160]	[67]
Cs <sup>(2)</sup>	GEA	1.5	37	130 <sup>(3)</sup>	U	KPA	600	206 <sup>(4)</sup>	87
Cu	ICP-AES	17	[1.2]	[>6](1)	V	ICP-AES	NMRQ	< 1.3	ND
Dy	ICP-AES	NMRQ	< 1.3	ND	Y	ICP-AES	NMRQ	< 1.3	ND
Eu	ICP-AES	NMRQ	< 2.5	ND	Zn	ICP-AES	17	< 1.3	< 2.8
Fe	ICP-AES	150	[6.1]	[29]	Zr	ICP-AES	NMRQ	< 1.3	ND
K	ICP-AES	75	[57]	[0.03]	TOC	Hot Pers.	1500	< 30	< 0.3
La	ICP-AES	35	< 1.3	ND	TOC	Furnace	1500	180	0.4

Analyte	Method	MRQ μg/mL	Cs eluate µg/mL	% recovered	Analyte	Method	MRQ μg/mL	Cs eluate µg/mL	% recovered
Li	ICP-AES	NMRQ	< 0.75	ND	Br <sup>-</sup>	IC	NMRQ	< 63	ND
Mg	ICP-AES	300	< 2.5	ND	Cl <sup>-</sup>	IC	3	< 63	< 0.7
Mn	ICP-AES	150	< 1.3	ND	F <sup>-</sup>	IC	150	< 63	< 0.5
Mo	ICP-AES	150	< 1.3	ND	$NO_2^-$	IC	3000	< 125	< 0.1
Na	ICP-AES	75	912	0.14	$NO_3$	IC	3000	29250	NA
Nd	ICP-AES	NMRQ	< 2.5	ND	$PO_4^{-3}$	IC	2500	< 125	ND
Ni	ICP-AES	30	[1.6]	[6]	$SO_4^{-2}$	IC	2300	< 125	< 0.5
P	ICP-AES	600	< 2.5	< 0.1	Oxalate	IC	NMRQ	< 125	ND
Pb	ICP-AES	300	[4.7]	[>6](1)	ОН	Titration	17	NM	NM
Pd	ICP-MS	NMRQ	5.9 E-3	NM	Wt% dried solids		0.1	0.34%	NA
Pr	ICP-MS	NMRQ	5.2 E-4	NM	Wt% oxides		NMRQ	0.40%	NA
Pt	ICP-MS	NMRQ	3.0E-2	NM	Density, g/mL		NMRQ	1.011	NA

ASR 6192, sample ID 01-1842

Notes: MRQ is minimum reportable quantity requested by Bechtel. NMRQ is no minimum reportable quantity requested. The overall error is estimated to be within +/-15%. Values in brackets are within 10-times the detection limit and errors are likely to exceed +/-15%.

NA = not applicable

ND = analyte not detected in the feed or product

NM = analyte not measured in the eluate

HB = high preparative blank concentration, recovery cannot be calculated

- Analyte was not detected in the feed.
- Calculated based on <sup>137</sup>Cs concentration in the eluate and applied Cs isotopic distribution ratio (<sup>133</sup>Cs 60.6 wt%, <sup>135</sup>Cs 15.4 wt%, <sup>137</sup>Cs 24.0 wt%).
- Calculated based on Cs feed concentration of 4.89 µg/mL. If the feed Cs concentration is actually 6.04 μg/mL, then Cs recovery is calculated to be 105%. See footnote (1) of Table 2.3.
- (4) Total U measured by ICP-MS was in good agreement at 190 μg/mL.

**Table 3.5.** Radionuclides in the Lead Column Eluate Composite<sup>(1)</sup>

Analyte	Method	MRQ <sup>(2)</sup> μCi/mL	Cs eluate µCi/mL	Error %	% recovered	Analyte	Method	MRQ <sup>(2)</sup> μCi/mL	Cs eluate µCi/mL	Error %	% recovered
$^{3}H$	Radchem	NMRQ	1.72 E-4	10	NM	<sup>155</sup> Eu	GEA	9.00 E-2	<3 E-1	_	ND
<sup>14</sup> C	Radchem	NMRQ	2.4 E-5	30	NM	<sup>232</sup> Th	GEA	NMRQ	<2 E-1	_	NM
<sup>51</sup> Cr	GEA	NMRQ	< 2 E-2	-	NM	$^{233}U$	ICP-MS	NMRQ	1.78 E-4	1	NM
<sup>54</sup> Mn	GEA	NMRQ	<3 E-2	1	NM	$^{234}U$	ICP-MS	NMRQ	9.89 E-5	4	NM
<sup>59</sup> Fe	GEA	NMRQ	<3 E-2	ı	NM	<sup>235</sup> U	ICP-MS	NMRQ	3.53 E-6	1	NM
<sup>60</sup> Co	GEA	NMRQ	<5 E-3	ı	<34	<sup>236</sup> U	ICP-MS	NMRQ	7.17 E-6	2	NM
<sup>63</sup> Ni	Radchem	NMRQ	8.83 E-4	5	NM	<sup>238</sup> U	ICP-MS	NMRQ	6.31 E-5	0.5	NM
<sup>79</sup> Se	Radchem	NMRQ	3.52 E-6	25	NM	<sup>236</sup> Pu	Radchem	NMRQ	< 4 E-7	_	NM
<sup>88</sup> Y	GEA	NMRQ	<2 E-2	1	NM	<sup>237</sup> Np	ICP-MS	NMRQ	6.25 E-7	13	ND
<sup>90</sup> Sr	Radchem	1.50 E-1	2.95 E-2	3	7	<sup>238</sup> Pu	Radchem	NMRQ	6.00 E-5	4	74
95Nb	GEA	NMRQ	<2 E-2	_	NM	<sup>239+240</sup> Pu	Radchem	NMRQ	4.63 E-4	2	72
<sup>99</sup> Tc	ICP-MS	3.00 E-3	1.36 E-4	10	0.06	<sup>239</sup> Pu	ICP-MS	NMRQ	7.64 E-4	3	NM
<sup>103</sup> Ru	GEA	NMRQ	<2 E-1	ı	NM	<sup>240</sup> Pu	ICP-MS	NMRQ	1.34 E-4	13	NM
<sup>106</sup> Ru	GEA	NMRQ	<7 E-1	ı	NM	<sup>241</sup> Pu	Radchem	NMRQ	3.06 E-3	8	NM
<sup>113</sup> Sn	GEA	NMRQ	<2 E-1	-	NM	<sup>241</sup> Am	Radchem	7.20E-04	3.04 E-5	3	3.8
<sup>125</sup> Sb	GEA	NMRQ	<4 E-1	ı	NM	<sup>241</sup> Pu, <sup>241</sup> Am	ICP-MS	NMRQ	<1.6 E-4 με	g/mL <sup>(5)</sup>	NM
126Sn/Sb	GEA	NMRQ	2.69 E-1	19	NM	<sup>242</sup> Am	Radchem	NMRQ	< 3 E-7	_	NM
$^{129}I$	ICP-MS	NMRQ	< 6 E-6	1	NM	<sup>242</sup> Cm	Radchem	NMRQ	<2 E-7	_	ND
<sup>134</sup> Cs	GEA	NMRQ	1.56 E-1	6	NM	<sup>242</sup> Pu	ICP-MS	NMRQ	<1.6 E-6	_	NM
<sup>137</sup> Cs	GEA	5.00 E-2	7.65 E+2	2	104	<sup>243+244</sup> Cm	Radchem	NMRQ	7.30 E-6	6	57
<sup>144</sup> Ce	GEA	NMRQ	<6 E-1	-	NM	Alpha sum <sup>(4)</sup>	Radchem	NMRQ	5.60 E-4	2	36
<sup>151</sup> Sm <sup>(3)</sup>	Radchem	NMRQ	2.42 E-4	6	NM	Total alpha	Radchem	2.30E-01	9.13 E-4	11	
<sup>152</sup> Eu	GEA	NMRQ	<1 E-2	ı	NM	Total beta	Radchem	NMRQ	7.97 E+2	4	NM

Analyte	Method	MRQ <sup>(2)</sup> μCi/mL	Cs eluate μCi/mL	Error %	% recovered	Analyte	Method	MRQ <sup>(2)</sup> μCi/mL	Cs eluate µCi/mL	Error %	% recovered
<sup>154</sup> Eu	GEA	2.00 E-3	<3 E-2	-	ND						

#### Notes:

NM = not measured in the feed

ND = not detected in feed or eluate

- (1) ASR 6192, Sample 01-1842
   (2) MRQ is minimum reportable quantity requested by Bechtel. NMRQ is no minimum reportable quantity requested.
   (3) The <sup>151</sup>Sm is an upper bound because other beta-emitters interfered with the analysis.
   (4) The alpha sum (total of individually-measured Pu, Am, and Cm alpha emitters) provides the best estimate of the total alpha activity in the sample.
- (5) The AMU-241 is reported in units of  $\mu$ g/mL. Applying the specific activity of <sup>241</sup>Pu, sample activity is <1.6E-2  $\mu$ Ci/mL; applying the specific activity of <sup>241</sup>Am, sample activity is <5.4E-4  $\mu$ Ci/mL.

### 3.2.5 Regeneration

The lead column was regenerated with 50 mL 0.25 M NaOH, collected in 1 composite fraction. The composition of this regeneration effluent solution is shown in Table 3.6. Sodium was the only cation found above the blank concentration. The charge balance between Na and OH are not equal. The counter-ion for the extra Na is predicted to be residual NO<sub>3</sub> from the elution step.

Analyte	Concentration, µg/mL	Concentration, M				
Na <sup>+</sup>	1020	0.044				
K <sup>+</sup>	< 40	< 1E-3				
OH-	500	0.030				
Total Cs <sup>(1)</sup>	2.8 E-3	2.1 E-8				
<sup>137</sup> Cs	6.69 E-4	5.85 E-2 μCi/mL				
Density, g/mL	1.006					
(1) Total Cs is calcu	(1) Total Cs is calculated from the isotopic ratio and measured <sup>137</sup> Cs concentration					

**Table 3.6.** Composition of Regeneration Solution

## 3.2.6 Activity Balance for <sup>137</sup>Cs

An activity balance for  $^{137}$ Cs was completed to compare the  $^{137}$ Cs recovered in various process streams to the  $^{137}$ Cs present in the feed sample (Table 3.7). As expected, virtually all  $^{137}$ Cs was found in the eluate, recovering >99 % of the  $^{137}$ Cs present in the initial AP-101DF feed. The lag column remained loaded with 35  $\mu$ Ci  $^{137}$ Cs based on the integration of the lead column breakthrough curve.

Solution	<sup>137</sup> Cs, μCi	137Cs in Feed Sample, %		
Feed Sample	1.50 E+5	1.00 E+2		
Effluent	1.3 E-1	8.4 E-5		
Load samples	1.2 E+0	8.2 E-4		
Feed displacement	2.6 E-1	1.7 E-4		
DI Water Rinse	2.1 E-1	1.4 E-4		
Column #1 Eluate	1.56 E+5	1.04 E+2		
Column #1 DI water rinse	7.6 E+0	5.1 E-3		
Column #1 regeneration	3.0 E+0	2.0 E-3		
Lag column Cs loading	3.5 E+1	2.3 E-2		
Total <sup>137</sup> Cs Recovery	1.56 E+5	1.04 E+2		

**Table 3.7.** Activity Balance for <sup>137</sup>Cs

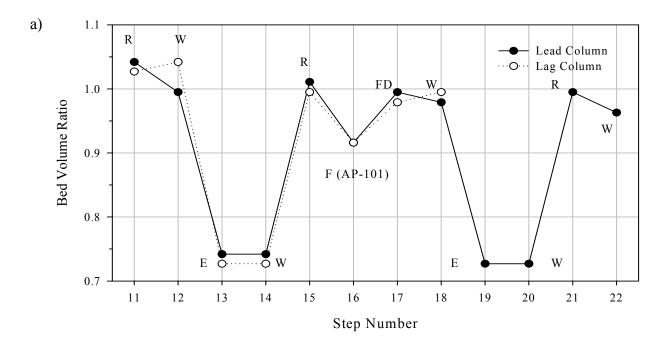
### 3.2.7 SL-644 Resin Volume Changes

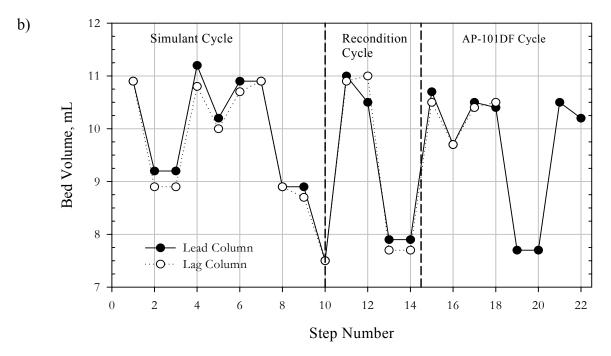
The SL-644 resin is known to change in volume as a function of the solution pH and ionic strength (Hassan et al., 1999). The resin BV change history is shown in Table 3.8. The volume contraction after each subsequent 0.5 M HNO<sub>3</sub> step became more pronounced with cycling. The first volume contraction stabilized at 9.2 mL; the final measured volume contraction stabilized at 7.7 mL. The variation in BV as a function of the process steps for both columns is shown in Figure 3.5 a and b. In Figure 3.5a, the BVs are normalized to the volume in the 0.25 M NaOH regeneration condition just prior to AP-101DF loading. Two process cycles are shown with the first cycle consisting of the bed conditioning steps and the second cycle consisting of the actual process test. Each process step is denoted with a number given in Table 3.8. In Figure 3.5b, the observed volume changes show clearly the greater contraction and slightly reduced expansion. Fluidizing the bed in the H-form resulted in tighter resin packing, yet the subsequent expansion in the Na-form appears to have been unaffected in that the regeneration volume is equivalent to that obtained prior to fluidization. Overall, there appears to be a general decline in resin volume in both the H-form and the Na-form. The change in resin volume may be caused in part by small losses of the resin during processing. Resin volume changes may also be attributed to compaction while swelling during the regeneration step.

**Table 3.8.** SL-644 Bed Volume Changes

		Process	BV 010319SMC-IV-73 212-425 μm PSD			
Feed	Symbol	Step	Lead column, mL(1)	Lag column, mL <sup>(2)</sup>		
Initial packing	P	1	10.9	10.9		
0.5 M HNO <sub>3</sub>	Е	2	9.2	8.9		
DI water	W	3	9.2	8.9		
0.25 M NaOH	R	4	11.2	10.8		
AW-101 simulant	F	5	10.2	10.0		
0.1 M NaOH	FD	6	10.9	10.7		
DI water	W	7	10.9	10.9		
0.5 M HNO <sub>3</sub>	Е	8	8.9	8.9		
DI water	W	9	8.9	8.7		
Re-fluidize bed	RP	10	7.5	7.5		
0.25 M NaOH	R	11	11.0	10.9		
DI water	W	12	10.5	11.0		
0.5 M HNO <sub>3</sub>	Е	13	7.9	7.7		
DI water	W	14	7.9	7.7		
0.25 M NaOH	R	15	10.7	10.5		
AP-101DF	F	16	9.7	9.7		
0.1 M NaOH	FD	17	10.5	10.4		
DI water	W	18	10.4	10.5		
0.5 M HNO <sub>3</sub>	Е	19	7.7	_		
DI water	W	20	7.7	_		
0.25 M NaOH	R	21	10.5			
DI water	W	22	10.2	_		

Calculated resin mass after washing, dry Na-form: 2.46 g.
 Calculated resin mass after washing, dry Na-form: 2.48 g. Note: The inside diameter of each column was 1.46 cm.





**Figure 3.5.** (a and b) Comparison of Bed Volumes of the Lead and Lag Columns (SL-644 Batch 010319SMC-IV-73, 212- to 425-μm)

- a) Relative BV for reported testing cycle; baseline BV = 10.6 mL in the 0.25 M NaOH Regeneration Condition.
- b) Actual BVs for all tested cycles, including initial simulant run.

### 4.0 Conclusions

The objectives of the testing were met.

Batch distribution values were developed as a function of Na/Cs mole ratios for both SL-644 and IE-911 in AP-101DF.

- The SL-644 equilibrium data resulted in a K<sub>d</sub> value of 690 mL/g, corresponding to a predicted Cs λ of 160 BVs (0.25 M NaOH condition), at a Na/Cs mole ratio of 1.4 E+5, 25°C.
- The IE-911 CST equilibrium data resulted in a  $K_d$  value of 710 mL/g, corresponding to a Cs  $\lambda$  value of 740 BVs, at a Na/Cs mole ratio of 1.4E+5, 25 °C.

Cs decontamination from AP-101DF (Envelope A) was successfully demonstrated.

An overall DF of 7.8 E+5 was obtained after processing 112 BVs providing a Cs-decontaminated effluent with a <sup>137</sup>Cs concentration of 1.1 E-4 μCi/mL. This represents 0.09% of the contract limit of 1.25 E-1 μCi/mL in the treated effluent (based on 14 wt% waste Na<sub>2</sub>O loading). <sup>15</sup>

Cs load and elution breakthrough profiles were developed.

- The estimated 50% Cs breakthrough for the lead ion exchange column, extrapolated from 0.2% C/C<sub>o</sub> to 50% C/C<sub>o</sub>, was about 190 BVs (0.25 M NaOH regeneration condition). This is a huge extrapolation and as such is inherently uncertain, however it is nearly equivalent to the predicted 160 BVs based on batch-contact studies.
- The Cs-loaded lead column was efficiently eluted with 0.5 M HNO<sub>3</sub>. Over 94% of the  $^{137}$ Cs was eluted from the column in 2.5 BVs of eluate. A total of 12 BVs eluate were required to reach the elution end point of  $C/C_o = 0.01$ . The peak  $C/C_o$  value for  $^{137}$ Cs was 64 for the lead column. The activity balance accounted for 104% of the feed  $^{137}$ Cs, which is essentially equivalent to 100% within experimental error.

The Cs eluate solution was composited and characterized in preparation for subsequent evaporation and HLW vitrification testing.

• A high percentage of the U (87%) and Pu (72%) was exchanged onto the resin and recovered in the acid eluate.

4.1

 $<sup>^{15}</sup>$  The effluent  $^{137}Cs$  concentration was 0.13% of the contract limit of 0.0874  $\mu Ci/mL$  in the treated effluent based on the maximum waste loading of 20 wt% waste Na<sub>2</sub>O.

The effectiveness of all SL-644 ion exchange process steps was demonstrated.

- The SL-644 was adequately regenerated with 4.7 BVs or 2.2 AVs of 0.25 M NaOH. The average  $^{137}$ Cs concentration in the regeneration solution was 5.8 E-2  $\mu$ Ci/mL, equivalent to C/C<sub>o</sub> of 4.6 E-4.
- An activity balance for <sup>137</sup>Cs indicated 104% of the <sup>137</sup>Cs present in the feed sample was accounted for in the samples and process streams (mostly in the eluate) indicative of good experimental integrity.
- No fouling, or other off-normal conditions, of the resin bed was observed.

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## Appendix A

**Calculations** 

#### Cs-137 Contractual Limit in AP-101 Env. A Vit feed

#### Assumptions

- 1) Concentration of Na<sub>2</sub>O in Env. A glass = 14% (=14 g Na<sub>2</sub>O / 100 g glass)
- 2) For maximum <sup>137</sup>Cs concentration in glass assume all Na comes from the feed. If some Na is added to Vit feed, multiply the maximum <sup>137</sup>Cs value determined below by ratio of total Na:feed Na.
- 3) Glass density =  $2.66 \text{ MT/m}^3$  (=2.66 g/mL)
- 4) Maximum Cs-137 in glass =  $0.3 \text{ Ci/m}^3$  (= 0.3 Ci / 1E+6 mL = 3E-7 Ci/mL)
- 5) AP-101DF actual waste and AW-101 simulant Na concentration = 5 M
- 6) AP-101DF actual waste  $^{137}$ Cs concentration = 126  $\mu$ Ci / mL

#### Na Loading in Glass

14 g Na<sub>2</sub>O / 100g glass \* 1 mole Na<sub>2</sub>O / 62 g Na<sub>2</sub>O) \* (2 mole Na/ mole Na<sub>2</sub>O)\* (23 g Na / mole Na) \* (2.66 g glass / mL glass) = 0.276 g Na / mL glass

Maximum <sup>137</sup>Cs:Na in glass

 $(3.0\text{E-7 Ci}^{137}\text{Cs} / \text{mL glass}) / (0.276 \text{ g Na} / \text{mL glass}) = 1.09 \text{ E-6 Ci}^{137}\text{Cs} / \text{g Na}$ 

 $(1.09 \text{ E-6 Ci}^{137}\text{Cs}/\text{g Na}) * (23 \text{ g Na}/\text{mole}) = 2.50\text{E-5 Ci}^{137}\text{Cs}/\text{mole Na}$ 

Maximum <sup>137</sup>Cs:Na in feed

(2.5E-5 Ci 
$$^{137}$$
Cs / mole Na) \* (5 mole Na / L feed) = 1.25 E-4 Ci  $^{137}$ Cs / L = 125  $\mu$ Ci  $^{137}$ Cs / L = 0.125  $\mu$ Ci  $^{137}$ Cs / mL

### AP-101DF actual waste Cs fraction remaining (C/Co) Contractual Limit

$$(0.125 \mu \text{Ci}^{137}\text{Cs/mL}) / (126 \mu \text{Ci}^{137}\text{Cs/mL})$$
 = 9.92E-4 C/C<sub>o</sub> = 0.0992 % C/C<sub>o</sub>

### Decontamination Factor (DF) Contract Limit

$$1/(9.92E-4 C/C_0) = 1008 C_0/C$$

# Appendix B

**Sample Identification** 

Table B.1. Sample Identification

Sample Description	ASR	RPL ID	Sample Identification	<b>Extended Sample Description</b>	
Batch Contact Testing					
First batch contact with	6048	01-0575 01-0576	AP101-644a and duplicate	Batch contact with SL-644, no spike	
SL-644 981112YK-N3-16/18		01-0577 01-0578	AP101-S1-644a and duplicate	Batch contact with SL-644, 1E-3 M Cs spike	
		01-0579 01-0580	AP101-S2-644a and duplicate	Batch contact with SL-6444, 5E-3 M Cs spike	
Batch contact with IE-911		01-0581 01-0582	AP101-CSTa and duplicate	Batch contact with IE-911 no spike	
		01-0583 01-0584	AP101-S1-CSTa and duplicate	Batch contact with IE-911, 1E-3M Cs spike	
		01-0585 01-0586	AP101-S2-CSTa and duplicate	Batch contact with IE-911, 5E-3M Cs spike	
Controls		01-0587	AP101-S1a	Batch contact control (no exchanger) 1E-3M Cs	
		01-0588	AP101-S2a	Batch contact control (no exchanger) 5E-3M Cs	
		01-0589	AP101-Ca	Batch contact control (no exchanger), unspiked	
Second batch contact with	6130	01-133601-1337	AP101-S3-644-F/A and duplicate	Batch contact samples with SL-644, 1E-3M Cs	
SL-644 010319SMC-IV-73		01-1338	AP101- S3C-F/A	Batch contact control (no exchanger) Spike 3	
		01-1339 01-1340	AP101-S2a and S1a	Re-analysis of controls, Spike 1 and Spike 2	
Column Run					
Initial Feed Sample, AP-					
101DF	6031	01-0520	AP-101DF	Initial feed sample	
Load - Elution samples	6097	01-0780	AP-101-F0	Initial feed sample	
		01-0781 01-0785	AP-101L-F3 through F-13	Lead column load samples	
		01-0786 01-0788	AP-101P-F4 through F-13	Lag column load samples	
		01-0789 01-0791	AP101-Fcomp2-A through Fcomp4-A	Composite effluent samples	
		01-0792 01-0794	AP101-FD2-A through FD6-A	Feed displacement samples	
		01-0795 01-0796	AP101-DI2-A and DI5-A	Water rinse samples	
		01-0797 01-0809	AP101L-E1-DA	Lead column elution analytical samples	
		01-0810	AP101L-RGN-A	Regeneration solution analytical sample	
AP-101DF effluent composite	6121	01-1203	AP1-Tc-0-C	AP-101DF Cs IX effluent composite	
AP-101 Cs eluent	6192	01-1842	AP101L-E-Comp	AP-101DF Cs eluant	

Notes: 1) The suffix letter "a" after the sample ID indicates the sample was loaded out of the hot cell into a clean analytical vial; the suffix letter "D" after the sample ID indicates the sample was diluted prior to submission for analysis. Unidentified samples are associated with other tests unrelated to this work.

# Appendix C

**AP-101DF Batch Contact Analytical Results** 

## Appendix D

**AP-101DF Column Processing Calculations and Results** 

# Appendix E

**AP-101DF Column Processing Analytical Results** 

## **Appendix F**

## **Cesium Ion Exchange and Batch Contacts Testing Personnel**

### **Cognizant Scientists**

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- D. L. Blanchard
- S. T. Arm
- B. M. Rapko

### **Hot Cell Technicians**

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- M. A. Mann

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